Montana Department of Environmental Quality

Sample Collection and Laboratory Analysis of Chlorophyll-*a* Standard Operation Procedure

February 23, 2011

Approvals:

_____

Rosie Sada, Monitoring and Assessment Manager, MDEQ/WQPB   Date

_____

Mindy McCarthy, QA Officer, MDEQ/WQPB   Date

_____

Michael Suplee, Ph. D., Water Quality Standards, MDEQ/WQPB   Date

REVISION HISTORY

Revision No.	Date	Modified By	Sections Modified	Description of Changes
1				
2				
3	2006	M. Bostrom	Multiple	Required the use of 11-transect system for assessing wadeable stream sites; required use of hoop, template or core as base Chl*a* field methods.
4	6/05/2008	C. Shirley	Section 6	Details on Chl*a* analysis coefficients for different extraction solutes. Section added on compositing individual Chl*a* samples, by method.
5	2/14/2011	M. Suplee	Mainly 1, 2; also 3.1.3 and 6.2.4	Clarified field sampling methods. Added Appendix A, describing overall statistical confidence in average algae density (mg/m^2) derived using the SOP method. Added Appendix B (updated Aquatic Plant Tracking Form).

TABLE OF CONTENTS

LIST OF TABLES

LIST OF FIGURES

ACRONYMS

ACF	Absorbance Correction Factors
APHA	American Public Health Association
ARM	Administrative Rules of Montana
DEQ	Department of Environmental Quality (Montana)
EDD	Electronic Data Deliverable
EMAP	EPA Environmental Monitoring and Assessment Program
EPA	Environmental Protection Agency (US)
FERC	Federal Energy Regulation Commission
GF/F	Glass Fiber Filters
GPS	Global Positioning System
HPLC	High-performance liquid chromatography
MDEQ	Montana Department of Environmental Quality
QAPP	Quality Assurance Project Plan
SAP	Sampling and Analysis Plan
SOP	Standard Operating Procedures
STORET	EPA STOrage and RETrieval database
SVF/COC	Site Visit Form/ Chain of Custody Form
TMDL	Total Maximum Daily Load
WQPB	Water Quality Planning Bureau (DEQ)
WQX	EPA's Water Quality Exchange System

1.0 CHLOROPHYLL-*A*

Chlorophyll-*a* (Chl*a*) is measured as a means of estimating algae (periphyton or phytoplankton) biomass in a body of water. It is expressed as either mass/area for periphyton (mg/m^2), or as mass/volume for plankton species (µg/L). Heavy growths of algae generally indicate inferior water quality.

Excess algae growth may clog water filters and irrigation equipment, cause taste and odor problems in water supplies, reduce dissolved oxygen levels, interfere with fish spawning, degrade macroinvertebrate habitat, trap sediment, deflect stream flows, and impair the overall aesthetics and recreational value of a stream.

1.1 SCOPE AND APPLICABILITY

This method is developed for use in water quality assessment decision-making. The sampling methods described herein are largely for wadeable streams and rivers. The phytoplankton sampling procedure may be used in low-flow conditions, including disconnected series of pools, and in lakes and reservoirs. These sampling methods are designed to produce a quantitative measure of algae growth by relating the total mass of Chl*a* pigment to a known area or volume.

1.2 SAMPLING DESIGN CONSIDERATIONS

1.2.1 Index period
Periphyton growth is controlled by season, nutrient concentrations, velocity of the current, days of accrual, shading, water temperature, and other factors. Because of this, sampling designs using Chl*a* must be inclusive of the times when stable flows have been achieved, as well as times when diversity and standing crop are peaking. Intensive sampling may include multiple visits to show the water body's baseline condition, period of high growth potential, and subsequent return to baseline conditions. The summer and early fall period of July 1st to September 30th is generally the time of maximum growth potential in western Montana (mountainous region). A somewhat longer sampling index period (June 16th to September 30th) is recommended for some plains ecoregions (Suplee et al., 2008).

1.2.2 Recent conditions
Sampling events planned in advance must consider the possibility that current or recent weather patterns could influence the sampling outcome. An example of this is recent rainfall that has significantly increased the flow, scouring the substrate. If the water body has had recent significant rainfall or is currently experiencing a significant rainfall event, consider the effect of scouring and reschedule sampling event, as needed.

1.2.3 Site Locations
Selection of sampling locations depends largely on the data quality objectives (DQOs) of the water quality study. The study design necessary to satisfy these DQOs must be documented in a project plan (QAPP, SAP, or equivalent documentation). The project plan should have sufficient detail to allow minor adjustments (of pre-selected sites) in the field due to unforeseen events such as site inaccessibility.

If sampling locations are to be determined in the field, field guidance should include a rationale for site selection. This is critical when different sampling crews select representative locations based on professional judgment.

1.2.4 Geo-locating site

The first measurement collected is a geo-reference for the study site. The location to be recorded is the F transect (layout of the sampling frame is discussed in **Section 1.2.5**). This is consistent with the approach used in EPA's Environmental Monitoring and Assessment Program (EMAP) studies (Lazorchak et al., 1998) as well as DEQ's Reference Project. The "F" location will be used for geo-referencing the site for EPA's STORET/WQX database. Once the site is located in the field, use a GPS receiver and record the latitude/longitude on the Site Visit Form. Always use Datum NAD83 coordinates.

1.2.5 Sampling Frame

A sampling frame will consist of 11 transects, with the total frame length equal to 40 times the average wetted width at the F site (**Figure 1.0**). Small streams must use 40 wetted widths or a minimum of 150 meters, whichever is larger. The procedure for determining the average wetted width at the F site is: measure the wetted width at 5 places around the F site (2 upstream, 2 downstream, and 1 at the F site), average the five readings, and round to the nearest 1 m. Since rivers can be quite wide, 40 wetted widths could entail kilometers, which is unmanageable. Therefore, if the sampling frame would exceed approximately 500 meters then it is acceptable to switch to the large river method. This entails taking transect samples across the river at a defined point ("F" in **Figure 1.0**)[1].

Samples taken at each location within the frame are *single* collections using the appropriate collection technique for the substrate encountered (listed in **Section 2** - Sample Collection Methods). The starting point right (R), left (L), or center (C) should be randomly selected at the most downstream transect (transect A). Place the remaining sampling locations progressing upstream following the R, L, C pattern. (**Note**: If a duplicate sampling event is desired, repeat the entire process but commence the duplicate's pattern at transect A at one of the two remaining transect starting points [e.g., if R was used for the first sampling, use L or C]. Follow the pattern upstream accordingly.)

The overall statistical confidence in benthic Chl*a* averages derived from an 11-replicate sampling frame as outlined above is provided in **Appendix A**. It has been found that for a typical wadeable stream benthic Chl*a* sampling event that has followed this SOP, DEQ is confident that <u>at least</u> 80% of the time the <u>measured</u> Chl*a* average will be within ± 30% of the <u>true</u> average.

---

[1] Other approaches for sampling rivers are being evaluated (e.g., increasing the number of samples in the wadeable region of the river while reducing the number in the deeper water). Future versions of this SOP will likely include other alternatives.

o - designates sample locations, wadeable streams and rivers
® - designates sample locations, large wadeable rivers
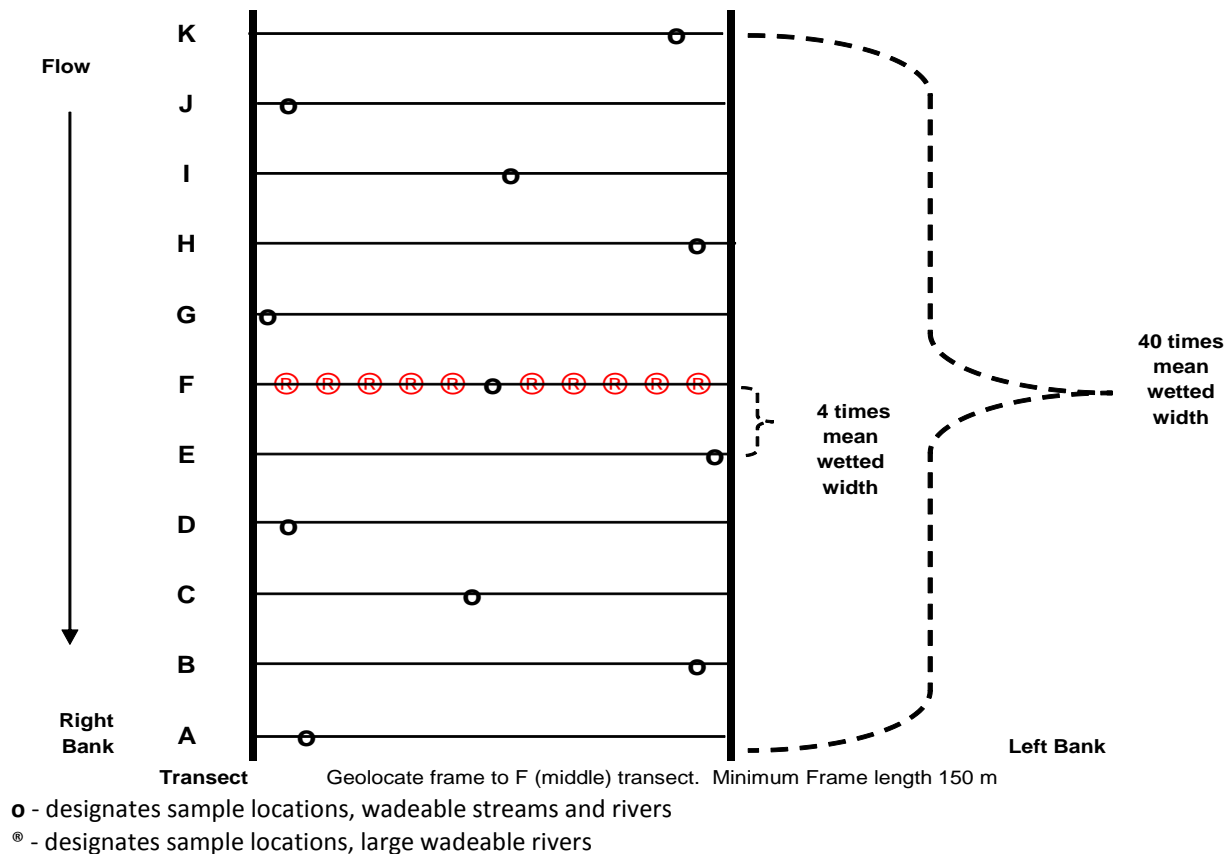
Figure 1-0. Diagram of Chl*a* sampling frame (wadeable streams & rivers).

Note: If the sampling frame exceeds about 500 m, switch to large river method (eleven evenly spaced samples taken <u>across</u> the river/stream at site F.)

1.2.6 Sampling Quality Control

The appropriate Quality Control samples to assess field collection activities must be designated in the project planning documents (QAPP, SAP). Because the designated sampling frame is a multi–transect sampling, information about the variability among measurements is inherent to the collection design. Therefore, duplicate samples do not *generally* need to be collected unless project DQOs require a high degree of defensibility. Documentation of the approach intended to be used to evaluate the results should be described in the quality control section of the project planning document(s).

1.2.7 Data Review and Evaluation

Data interpretation is based upon the arithmetic mean of the eleven individual results. If samples are composited according to type (core, hoop, template), a weighted average Chl*a* value (**Section 5.2.4**) is obtained based on the number of transects per collection method.

2.0 SAMPLE COLLECTION METHODS

Periphyton standing crop may be quantified by measuring the amount of accrual on natural substrates at the study site. The sampling of artificial substrates is not recommended.

There are three methods for collecting attached algae (periphyton) from streams and rivers – the hoop, the core, and the template. A single sampling using one of the listed collection techniques is performed at each transect. The substrate and conditions encountered at the sampling locale on the transect determine the collection technique. At each of the 11 sampling locales, the algae sample collected should represent conditions prevalent in an approximately 1 m^2 area around the transect. For example, if the sample is to be collected at transect D, Right (**Figure 1.0**), the sampler should observe the algae conditions that prevail from the right wetted edge to 1m out along and a half a meter up- and downstream of the transect line. The sampler then selects the appropriate sampling method (hoop, core, template) and samples the most representative point. For Center samples, observe 0.5 m on four sides of the channel centerpoint (upstream, downstream, towards river R, towards river L) and then sample the representative point. Using the form in **Appendix B**, track the transect sampling point (R, L, C), method used, approximate proportion of macrophytes vs. algae (for hoops), and the presence of moss.

The 11 transect sampling frame (**Section 1.2.5**) is required for all collections. It has been demonstrated that this design will generally encounter one to three extreme values; thus the relatively high coefficient of variation (73%) typically associated with the 11 samples (**Appendix A**). Analyzing each of the 11 samples separately allows the assessor to understand the patchiness of algal growth and to calculate the replicates' coefficient of variation (CV) for the sampling event. However, whether the data were each analyzed separately, or composited (discussed below), the sampling event's average is what is typically compared to recommended or adopted Chl*a* stream-impairment endpoints[2].

If the water is too deep (greater than 2-3 feet) in the location you are sampling (R, L, or C), adjust the sample location to the subsequent sampling locale in the pattern. Resume the normal sampling pattern (as originated at transect A) as soon as practicable. The form in Appendix B will aide in this.

Sample Compositing: Sample compositing may be used to reduce the costs associated with the 11 samples collected as part of the sampling frame described in **Section 1.2.5**. Sample compositing will, in effect, return results of each collection method as a *mean* when the composite Chl*a* concentration is calculated to the sum of the areas collected. **THE LABORATORY WILL DO THE COMPOSITING**. In the field, the sampler should keep each of the 11 samples separate, but needs to check the "Composite at Lab" box on the DEQ's Site Visit Form/Chain of Custody (SVF/COC). (More on filling out this form is in **Section 3.2**.)

LABORATORY CONTRACTORS PLEASE NOTE THAT ONLY SAMPLES COLLECTED BY THE SAME FIELD TECHNIQUE CAN BE COMPOSITED! There are three sampling techniques used in this method (Hoop,

---

[2] It should be noted that the recreationally-based impairment threshold identified in Suplee et al. (2009) is based on the quantification of stream bottom algae using the same sample collection methods (template, hoop) presented in this SOP.

Core, and Template). Therefore, there may be up to *three* composite samples resulting from an 11-transect sampling event.

2.1 TEMPLATE SAMPLING METHOD

The template method is used for sampling transects with substrate dominated by small boulders, cobble, and gravel *without heavy filamentous growth*.

2.1.1 Method Summary

A template with a 12.5 cm2 area is placed on a rock with representative algae density which has been drawn from the designated locale at the transect line. The sampler should observe the algae density in a roughly 1 meter by 1 meter area centered on the sampling point and select a representative rock therein. The area within the template is then scraped into a container and filtered on site (0.70 um Glass Fiber Filters (GF/F)). At the laboratory, the filter will be extracted and the resulting extract measured for Chl*a*. The template can be made, for example, from a cut-off piece of PVC pipe (Schedule 40 - 1 1/2" nominal I.D.), which results in an internal area of approximately 12.5 cm2. Internal diameter of template should be checked and be within 3.93 to 4.05 cm (+/- 4% area error).

2.1.2 Sampling Equipment
- Waders or hip boots
- 50 cm3 centrifuge tube
- Knife for scraping rock
- Tooth brush for brushing rock
- Tap water in squirt bottle
- Shallow plastic pan to hold rock
- Hand pump vacuum with tubing
- Nalgene filtering unit
- GF/F filters (0.70 um)
- Tweezers or forceps
- Cooler with ice or dry ice (preferred)
- Aluminum foil
- Large and medium Ziploc bags
- Sharpie

2.1.3 Sample Collection

A representative rock is placed in the shallow pan and the template placed over the upper (light-facing) surface of the rock. All of the growing material within the template is scraped/scrubbed and placed in the pan. In certain cases the volume of algal material on the rock surface is small, therefore it is better to scrub the rock surface with a toothbrush and then rinse the rock surface and the toothbrush into the pan with a small volume of **tap** water (Note: Previous versions of this SOP listed de-ionized water. DO NOT USE de-ionized water as it may burst cells due to osmotic pressure differences.)

In some cases, the rocks are very small (smaller than template diameter, but still too large for core sampling). In this case, instead of using 1 representative rock, place several small representative rocks inside the template diameter, and follow the process as described in the above paragraphs, scrubbing the light-facing surfaces as best possible.

Field filtration (MUST BE PERFORMED IN THE FIELD): The rinse water/algae material that has been rinsed into the pan is field filtered on to a GF/F filter using a hand pump, and the filter placed in the centrifuge tube. Refer to **Section 2.4.3** for proper use of the Nalgene filtering unit. Minimize rinse water use to assure that all water will move through the GF/F filter.

Circumstances where rocks have very low, or high, levels of attached algae:

- Very Low Levels: In some instances levels of attached algae or so low that scrapings from a single template will result in very little material on the GF/F filter. Little or no color will be observed on the filter after filtration. To better assure that the sample is sufficient to achieve detectable levels, up to 3 templates from the same rock (or from other representative rocks in the observation locale) can be collected and all the scraped material is then captured on the same GF/F filter. Record the number of templates aggregated on the single GF/F filter on the Aquatic Plant Tracking Form (**Appendix B**).

- High Levels: If benthic algae density from a single template is so high that the GF/F filter clogs prior to all water passing through[3], the remaining algal material/water in the upper half of the Nalgene unit may be returned to the <u>clean</u> pan. Make sure the pan is clean and will therefore only contain material scraped from the single template sample in question. Then, load a 2nd GF/F filter on the Nalgene and filter the remaining water/algae material. Both filters are placed in the centrifuge tube together. Record the number of GF/F filters associated with the single template on the Aquatic Plant Tracking Form (**Appendix B**).

2.1.4 Sample Handling & Labeling

Place the GF/F filter(s) into an appropriate container (50 cm3 centrifuge tube). Sampling location is identified on an external label with the following information:

      a) Sample Type
      b) Activity ID
      c) Collection Date
      d) Waterbody Name
      e) Collector's Name

Fill out the outside label, place it on the centrifuge tube or container, and cover the label completely with a strip of clear tape. Wrap the tube with aluminum foil to exclude light, write the Activity ID on the lid with a SHARPIE. Place the centrifuge tube or container into a Ziplock plastic bag.

Immediately store the sample on dry ice (preferred), or ice, and away from light. Samples should be sent to the laboratory as soon as possible for Chl*a* analysis.

Record the transect number (A-K), collection position (Right, Left, Center), and the collection technique (C = Core, H = Hoop, T = Template) on the SVF/COC. If the corresponding surface area is different than the one indicated in this SOP, you must record it on the SVF/COC.

---

[3] This can also occur due to unusually high levels of fine sediment deposited on the rocks, intermixed with the algae.

2.2 HOOP METHOD

The hoop method is designed for transects *dominated by the presence of filamentous algae*, regardless of stream substrate.

2.2.1 Method Summary

The hoop collection method is a sample from a representative area where filamentous algae dominates the sampling locale, regardless of stream substrate. Upon collection, filamentous algae is physically separated from macrophytes (retain both portions if Ash Free Dry Weight is being analyzed). The filamentous algae sample is submitted to the laboratory for extraction and analysis of Chl*a*. The macrophytes portion may be analyzed for ash free dry weight to determine macrophyte biomass. The hoop can be made by wrapping a stiff wire around the bottom of a 5 gallon bucket. Check area by measuring hoop diameter and calculating for area of a circle (A=3.14*(D/2)2), adjust as necessary to arrive at an area of 710 cm2. The diameter of the hoop is approximately 30 cm.

Sampling Equipment:
- Waders or hip boots
- Large freezer storage bags
- Aluminum Foil
- Cooler with ice or dry ice (preferred)
- Metal hoop (30 cm diameter, 710 cm2 area)
- Scissors
- Tooth brush
- Knife for scraping rocks
- Sharpie

2.2.2 Sample Collection

Find designated location in the transect line. Within the 1.0 m2 sampling locale, locate a representative area to place the hoop. If a small number of macrophytes (< 5% by area) are present, they can be separated from the filamentous algae sample at the time of collection. If >5% macrophytes are present, collect the sample and perform the filamentous algae separation on the bank using a pan or bucket placed on a stable surface.

Place the hoop over the representative area. All the algal material within the hoop is collected (i.e. filamentous and non-filamentous). Scissors or a knife may be used to detach the filamentous algae from their substrate. Filaments originating inside the hoop that are streaming beyond it in the downstream direction are to be cut off along the lower edge of the hoop, and only the parts within the hoop are retained. Similarly, filaments originating upstream of the hoop which are streaming down into the hoop (and sometimes beyond it) are to be cut along the edge(s) of the hoop, retaining only the filament parts that originally fell within the hoop. Algae attached to rocks within the hoop are scraped into the Ziplock bag. Minimize the amount of water submitted by decantation (do not decant floating algae); gently squeezing the water out of filaments works well.

2.2.3 Sample Handling & Labeling

Place all filamentous algae collected at the site into a large Ziploc freezer bag. Sampling location is identified on an external label with the following information:
    a) Sample Type

   b) Activity ID
   c) Collection Date
   d) Waterbody Name
   e) Collector's Name

Fill out the outside label, place it on Ziploc bag, and cover the label completely with a strip of clear tape. Wrap the bag with aluminum foil leaving no space for light to enter. Place this wrapped bag into another large Ziploc bag and hand write the Activity ID on the outer bag with a **SHARPIE**. Immediately store the sample on dry ice (preferred), or ice, and away from light. Send samples to the laboratory as soon as possible for Chl*a* analysis.

Record the transect letter (A-K), collection position (Right, Left, Center), and the collection technique (C = Core, H = Hoop, T = Template) on the SVF/COC. If the corresponding surface area is different than the one indicated in this SOP, you must record it on the SVF/COC.

2.3 CORE METHOD

Method for transects dominated by silt-clay substrate without heavy filamentous growth. These substrate types are often dominated by varying thicknesses of microalgal mats, which can have Chl*a* levels comparable to those measured in templates and hoops.

2.3.1 Method Summary

The core collection method is a sample from a representative area where a silt-clay substrate dominates the sampling point on the transect, and luxuriant plant growth is not present. A core sample is taken from the substrate. The top 1cm of the core is sliced off the plug and placed in a centrifuge tube. The sample is sent to the laboratory for extraction & analysis.

2.3.2 Sampling Equipment

- Waders or hip boots
- Cut-off 60 ml syringes (5.6cm2)
- 50 cm3 centrifuge tubes
- Aluminum foil
- Small Ziploc bags
- Knife
- Sharpie

2.3.3 Sample Collection

A 5.6 cm^2 core sample is collected using a cut-off 60 cc syringe in a representative portion of the designated transect site location. (**Figure 2.1**)
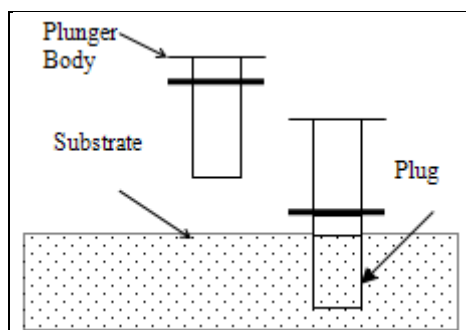
Figure 2-1 Illustration of Core Sample Collection Process.

Each core sample is taken by driving the 60 cc syringe into the substrate to a depth of 5-10 cm. The syringe plunger may have to be drawn up as the body of the syringe is pushed into the substrate, as the plunger may have too much friction within the barrel to rise up on its own.

The plug may be comprised of loose sediment that will fall out of the syringe. To minimize loss of a loose plug, the sampler can place his/her fingers over the end of the syringe as it is pulled out of the hole and up through the water column.

In deeper water, a broom handle can be duck-taped to the syringe to extend your reach.

Immediately invert the syringe containing the plug to prevent the plug from sliding out of the barrel. Extrude the core so the upper 1 cm of the core remains in the syringe (**Figure 2.2**). Slice off and discard the lower portion. Place the 1 cm portion in a 60 ml centrifuge tube.
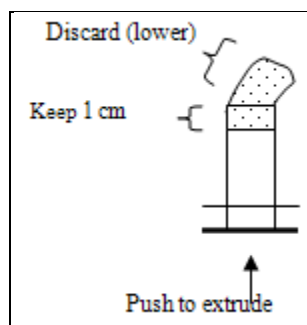
Figure 2-2 Extruding a Core Sample.

Important: Only the **upper 1 cm** of each core sample is placed in a centrifuge tube. Assure that all the material adhering to the rubber surface of the plunger-end is carefully collected, as most of the chlorophyll is located there.

2.3.4 Sample Handling & Labeling
Sampling location is identified on an external label with the following information:
    a) Sample Type
    b) Activity ID
    c) Collection Date
    d) Waterbody Name
    e) Collector's Name

Fill out the outside label, place it on the centrifuge tube, and cover the label completely with a strip of clear tape. Wrap it with aluminum foil to exclude light, write the Activity ID on the lid with a **SHARPIE**. Place the centrifuge tube or container into a self-sealing plastic bag.

Immediately store the sample on dry ice (preferred), or ice, and away from light. Samples should be sent to the laboratory as soon as possible for Chl*a* analysis. Record the transect letter (A-K), collection position (Right, Left, Center), and the collection technique (C = Core, H = Hoop, T = Template) on the SVF/COC. If the corresponding surface area is different than the one indicated in this SOP, you must record it on the SVF/COC.

2.4 PHYTOPLANKTON METHOD (CHL*A* IN WATER)

The phytoplankton method is the sampling method for determining Chl*a* in the water column. It is used for transects dominated by pools with green color (light green to dark green), and lakes.

2.4.1 Method Summary

This method uses a filtration apparatus to collect a sample. Since Chl*a* breaks down readily in sunlight, the use of a **dark Nalgene bottle is required** to minimize the exposure of the sample to sunlight[4]. The filter apparatus should be set up prior to sample collection to minimize time between sampling and filtration. The volume of water filtered must be recorded!

2.4.2 Sampling Equipment

- Snap-shut petri dish or 50 cm3 centrifuge tube
- 1 - hand pump vacuum with tubing
- Nalgene filtering unit
- GF/F filters (0.70 um)
- Tweezers or forceps
- Graduated cylinder (100-250 ml)
- Tap water in squeeze bottle
- 1 L (Dark) Nalgene Bottle
- Sharpie

2.4.3 Sample Collection

Filter apparatus setup
Using clean forceps, place a glass fiber filter (GF/F nominal pore size 0.7 um) on the filter holder. Use a small amount of tap water from a wash bottle to settle the filter. Rinse the sides of the filter funnel and filter with a small volume of tap water.

Sample collection and processing
Rinse a 1L dark Nalgene bottle 3 times with stream or lake water before collecting the sample.

---

[4] Sampling in lakes and large rivers often involves collecting various water samples at depth, which are then placed together in a single carboy. In such cases, assure that the carboy (which is usually made of plastic) is kept covered with a canvas tarp, or similar cover, to exclude light throughout the sampling process.

Grab a water sample from an undisturbed location using the 1L Nalgene bottle. Cap the bottle and invert the bottle 3 times to mix thoroughly. Rinse a 100-250 ml graduated cylinder three times with *tap water*. Measure 20 ml or more of sample water in the graduated cylinder and pour into the filter funnel, place the cap loosely on the filter funnel. Draw the sample through the filter using the vacuum hand pump. ***Note: To avoid rupture of fragile algal cells, do not exceed 9.0 inches Hg on the vacuum gauge.***

Keep track of the volume of sample water filtered! The volume of sample filtered may vary from 5 ml to 1000 ml or more. When filtration slows <u>and</u> the filter has developed a distinct green (or green-brown) color, sufficient sample has been filtered. Do not allow the filter to clog. If a filter completely clogs while water remains in the upper half of the apparatus, discard the filter and start again, using less water volume. (Note that this differs from the allowable approach for benthic algae templates presented in **Section 2.1.3.**)

After filtration is complete, unplug the hand pump, remove the filter funnel from the filter holder, and remove the filter with clean forceps. Avoid touching the colored portion of the filter. Fold the filter in half, with the colored side folded on itself. Place the folded filter paper inside the petri dish and snap it shut (or place the folded filter inside centrifuge tube).

2.4.4 Sample Handling & Labeling
Sampling location is identified on an external label with the following information:
      a) Sample Type
      b) Activity ID
      c) Collection Date
      d) Waterbody Name
      e) Collector's Name

Fill out the outside label, place it on the petri dish/centrifuge tube, and cover the label completely with a strip of clear tape. Wrap it with aluminum foil to exclude light, write the Activity ID on the foil with a **SHARPIE**. Place the petri dish/tube into a self-sealing plastic bag.

Immediately store the sample on dry ice. Always use dry ice for phytoplankton samples unless no other option is available; in such circumstances, regular ice may be used. Samples should be sent to the laboratory as soon as possible for Chl*a* analysis. Store the sample away from light.

3.0 RECORDING THE CHL*A* SAMPLING EVENT

The Chl*a* sampling event must be recorded including information on the location, collection methods used, and the area/volume collected. DEQ uses a SVF/COC to record this information in the field for later entry into the MT-eWQX database for submission into EPA's STORET database.

3.1 GEO LOCATION

As described in **Section 1.2.4** of this SOP, the F transect is used to geo-reference the sample site. Other locations may be recorded to document the extent of the sampling frame, however, these must be clearly distinguished so that the F middle transect is easily recognized. Regardless of how many locations are recorded, in the database only the F middle transect will be used to geo-locate the site.

3.2 RECORDING THE SAMPLING EVENT - DEQ SITE VISIT FORM/CHAIN-OF-CUSTODY (SVF/COC)

DEQ established a site visit form for recording monitoring sampling event metadata. This form was designed to geo-locate a sampling event to the single, downstream site in multi-parameter sampling events. Therefore, when using the DEQ site visit form, it is imperative that the geo-location of the F transect for a Chl*a* sampling event is distinguished from the latitude and longitude of the other measurements.

3.2.1 Filling Out the Form

Indicate a Chl*a* sample was taken by checking the "Chlorophyll a" box on the SVF/COC. Note the *Sample ID* on the form. Indicate the sample collection procedure and location used at each transect on the Site Visit Form using the following abbreviations. The first letter represents the sample type and the second letter represents the location on the transect.

First letter (Technique/Type)
- Template = T
- Hoop = H
- Core = C

Second letter (Position)
- Right = R
- Left = L
- Center = C

For example: Transect: A_C-R_, B_H-C__, C_C-L__, D_T-R__, Etc…

Record the geo-location of the F transect under Site Visit Comments. Record the mean wetted width from the F site (**Section 1.2.5** Sampling Frame) under Site Visit Comments. Include each sample on a separate line and include the area or volume collected so the laboratory may complete the calculations to area or volume. Phytoplankton samples must include documentation of volume filtered. ***The laboratory must not accept samples until the field crew provides these documents.***

3.2.2 Compositing

If compositing is desired, THIS FACT MUST BE NOTED ON THE SVF/COC by checking the box "Composite at Lab". Because Chl*a* readily breaks down in sunlight, samples must be composited in subdued light <u>at the laboratory</u> prior to processing. Samples collected by different techniques CANNOT be composited. However, method-specific composites may be made (i.e. Core Composite, Hoop Composite, and/or Template composite).

The sampler is responsible for providing the area, type (core, hoop, template), and the number of each type of each individual sample collected and recording this on the SVF/COC. The laboratory is responsible for multiplying the number of composites by the area of each sample for determining the denominator of the final result.

4.0 SAMPLE EXTRACTION

Sample extraction and spectrophotometric determination are to be performed in an analytical laboratory by a qualified laboratory technician or chemist. Sample extraction and determinative techniques described herein are *modified* from the procedure described in EPA 446.0 (Arar E.,1997). These modifications are:

- Use of the monochromatic equation for phaeopigment-corrected Chl*a* with the extraction solvent ethanol (Suplee et *al*., 2006), and
- Option for using high-performance liquid chromatography (HPLC).

The solvent purity and grade used for extraction can greatly influence the outcome of the analysis. Therefore, this procedure limits extraction solvent options to those listed in **Table 5-1**. If HPLC will be used, strict adherence to solvents and instrument conditions described in Standard Methods 10200H (5) must be maintained.

Caution: ALL CHLOROPHYLL WORK MUST BE PERFORMED IN SUBDUED LIGHT.

If processing must be delayed, hold solid and filter samples at -20°C and protect them from exposure to light. Solid and filter samples taken from water having a pH 7 or higher may be placed in airtight plastic freezer bags and stored frozen for 3 weeks. Samples from acidic water must be processed promptly to prevent Chl*a* degradation.

The four different sampling techniques result in different types of media which are submitted to the laboratory, and extractions to accommodate each media follow.

4.1 HOOP SAMPLES

This extraction must be performed in subdued light to minimize the degradation of Chl*a* pigment. A hoop sample consists of all filamentous algae collected in a 710 cm2 area. Samples are shipped to the laboratory in a Ziploc bag covered with aluminum foil, and with a second (outer) ziploc bag protecting the aluminum foil. Upon arrival at the laboratory, samples may be analyzed immediately or frozen until ready for analysis. Frozen samples must be thawed in subdued light at room temperature.

4.1.1 Sample Extraction
- If frozen, remove sample from freezer and allow it to thaw in subdued light at room temperature.
- Remove any excess water by straining or decanting. Excessive water interferes with the spectrophotometric measurement.
- Add enough solvent to the (inner) Ziplock bag to cover the sample. Record volume added. A minimum of 13 ml of solvent is needed for analysis. Additional solvent may be added, however the more dilute the sample becomes, the lower the instrument reading will be relative to the calibration range of the spectrophotometer - don't dilute into a non-detect at extraction.
- Samples may require grinding to break cell walls to release chlorophylls.
- Identify the sample with a label.
- Keep in the dark overnight.
- The next day, proceed with spectrophotometric analysis (**Section 5**).

4.2 CORE SAMPLES

This extraction must be performed in subdued light to minimize the degradation of Chl*a* pigment. A core sample is the *upper 1 cm* of a 5.6 cm2 sample of stream sediment. The sample will be returned from the field in a 50 cm3 centrifuge tube wrapped in aluminum foil. This foil wrapped tube should be in a protective (outer) Ziploc bag. Sample should arrive frozen and remain frozen until ready for extraction.

4.2.1 Sample Extraction
- When ready to extract, remove sample from freezer and allow it to thaw.
- Add enough solvent to the centrifuge tube to cover the sample. Record volume added. A minimum of 13 ml of solvent is needed for analysis. Additional solvent may be added, however the more /solvent added, the greater the dilution of pigments – *don't dilute into a non-detect.*
- Core samples can be very thick and compact. It may be necessary to mix the solvent and sample using a mechanical shaker.
- Keep in the dark overnight.
- The next day, proceed with spectrophotometric analysis (**Section 5**).

4.3 TEMPLATE SAMPLES

Template samples must arrive at the laboratory as a frozen filter sample. This extraction must be performed in subdued light to minimize the degradation of Chl*a* pigment.

Template samples are scrapings and rinsate from the rock, within the template area. The water and scrapings are then filtered. The filter is submitted to a laboratory who extracts the Chl*a* from the filter and analyzes the extract. Field crews normally have filtration equipment available and return samples on filters in centrifuge tubes. Regardless, the template area must be included on the SVF/COC prior to proceeding with the analysis because results are reported as mass/area. Filtered samples should be returned to the laboratory frozen and remain frozen until analysis.

4.3.1 Sample Extraction (note: first 3 bullets will normally have been done in *fi*eld)
- Assemble filtration apparatus and quantitatively transfer and filter entire sample through a 47 mm glass fiber filter with a nominal pore size of 0.7 um. (Whatman GF/F filters).
- Volume filtered is irrelevant because results will be related to template area.
- Place filter into a labeled centrifuge tube.
- Add a minimum of 13 ml. of solvent to centrifuge tube and keep in the dark overnight. Do not add more than 20 ml of solvent unless multiple filters were required for the sample. Make sure the entire filter is covered by solvent. Record volume of solvent added.
- Keep sample in the dark overnight.
- The next day, proceed with spectrophotometric analysis (**Section 5**).

4.4 PHYTOPLANKTON SAMPLES

This extraction must be performed in subdued light to minimize the degradation of Chl*a* pigment. Phytoplankton samples should arrive frozen at the laboratory as a filter in an aluminum-foil wrapped petri dish or similarly covered 50 cm3 centrifuge tube.

4.4.1 Filter Extraction

- When ready to extract, remove sample from freezer and allow it to thaw.
- Add enough solvent to the centrifuge tube (or sample beaker if sample was in a petri dish) to cover the sample. Record volume added. A minimum of 13 ml of solvent is needed for analysis. Additional solvent may be added, however the more solvent added, the greater the dilution of pigments – *don't dilute into a non-detect*.
- It may be necessary to mix the solvent and sample using a mechanical shaker.
- Keep in the dark overnight.
- The next day, proceed with spectrophotometric analysis (**Section 5**)

5.0 SPECTROPHOTOMETRIC DETERMINATION OF CHL*A*

The Spectrophotometric determination of Chl*a* is to be performed according to EPA Method 446.0 (Arar E.,1997) or Standard Methods 10200 H (APHA, 1998). Both of these methods include the monochromatic calculation required for spectrophotometric analysis with phaeophytin-*a* correction.

Each laboratory must have current standard operating procedures (SOPs) that describe their instruments, reagents, interferences, standards, instrument setup, calibration procedures, analytical procedures, quality control requirements, calculations, and reporting protocols. Except as provided below, these SOPs must describe a method in general accordance with the reference methods EPA 446.0 or Standard. Methods 10200H.

A reference sample must be run with each analysis to determine method bias at a 10% frequency. The reference may be purchased from Sigma Aldrich or any certified vendor. Acceptance limits are +/- 20% of True Value. Control charting of reference sample performance is suggested to better control method performance.

Chl*a* methods in general list Absorbance Correction Factors (ACF) or give values for k and A. It is preferred that the laboratory calculate the values of k and A from average values obtained by analyzing 20-30 aliquots of a reference material. This will allow the calculation of a method ACF specific to the laboratory's instrument and purity/grade of reagents used. If reference values (Table 1) from literature will be used rather than calculating its own ACF, the laboratory must demonstrate acceptable method performance in an initial method validation and re-establish this acceptable performance annually or as changes in instrument conditions or reagents require. Above all, recognize the potential for high or low bias to exist in this method and do not blindly follow published or literary values for the ACF without verification.

$$\text{Absorbance Correction Factor} = k \; X \; A$$

$$k \;\; = \;\; \frac{(664b/665a)}{(664b/665a)-1}$$

$$A \;\; = \;\; \frac{l \; (cm) \; X \; \text{Concentration (mg/l)}}{\text{Absorbance 664b}}$$

5.1 CORRECTION FOR PHAEOPHYTIN-A

Both reference methods (EPA 446.0 and Standard Methods 10200H) provide calculations for obtaining monochromatic (Chl*a* corrected for presence of phaeophytin-a) and trichromatic (Chl*a*,*b*,*c*) results. Montana law (ARM 17.30.602[4]) requires that Chl*a* water quality measurements be corrected for phaeophytin [pheophytin].)

Refer to Standard Methods 10200H(2) (APHA, 1998) or EPA Method 446.0 (Arar E.,1997), for instrument requirements, sample analysis requirements (calibrations, reagents, wavelengths, and calculations).

The calculations presented in the reference methods are for a phytoplankton (water) sample and can be applied directly for those samples.

Chl*a* mg/m3 = [(Absorbance Correction)((664b-750b)-(665a-750a))*V1/(V2*L)]

> Where: V1 = Volume of extract (L)
> V2 = Volume of sample (m3)
> a = after acidification
> b = before acidification
> L = Light path or width of cuvette , cm

The calculation for periphyton replaces area for volume.

Chl*a* mg/m2 = [(Absorbance Correction)((664b-750b)-(665a-750a))*V1/(A1*L)]

> Where: V1 = Volume of extract (L)
> A1 = Sample collection area (m2)
> a = after acidification
> b = before acidification
> L = Light path or width of cuvette , cm

The allowed solvents in its purity form are listed in **Table 5-1**.

Table 5-1. Approved Solvents and Absorbance Correction Factors.

Solvent Purity	Solvent	Absorption Peak Ratio(APR)	Specific Absorption Coefficients (E_{1cm})	A	K	Absorbance Correction = (A x K)
90%	Acetone (Ace)	1.7 [note 1]	89.0 L/(g*cm) [note 2]	11.0 [note 1]	2.43 [note 1]	26.7 [note 1,4]
95%	Ethanol (EtOH)	1.72 [note 2]	83.4 L/(g*cm) [note 2]	11.99 [note 3]	2.39 [note 3]	28.6 [note 3]

> 1. APHA, 1998
> 2. Values listed by Sartory (Sartory D.P. and Grobbelaar J.U.,1984)
> 3. Calculated from values listed by Sartory (Sartory D.P. and Grobbelaar J.U.,1984)
> 4. Significant figure error. Error carried forward.

5.2 CALCULATION TO AREA

In order to determine the density of periphyton algae by measuring Chl*a*, results obtained from the instrument (in mg) must be related to the area sampled rather than a volume of water. The area obtained from the three collection techniques varies. If the area information is not readily available on the SVF/COC, the laboratory must *not* begin the extraction and analysis until it is provided. Also, if compositing is used, the number of composites and total area sampled must be submitted on the SVF/COC.

5.2.1 Area of Hoops

A hoop has a standard area of 710 cm2. *Confirm area of hoop prior to use*. If compositing is used, the number of hoop samples composited must be confirmed from the SVF/COC. Calculate as follows:

$$\frac{710\ cm^2 \times "number of hoops"}{10,000\ cm^2/m^2} = X\ m^2$$

For example, for one hoop: $\dfrac{710\ cm^2 \times 1}{10,000\ cm^2/m^2} = 0.0710\ m^2$

5.2.2 Area of Cores

A 60 ml plastic syringe results in a core sample with a standard area of 5.6 cm^2. *Confirm area of syringe prior to use*. If compositing is used, the number of cores composited must be confirmed from the SVF/COC. Calculate as follows:

$$\frac{5.6\ cm^2 \times "number of cores"}{10,000\ cm^2/m^2} = X\ m^2$$

For example, for one core: $\dfrac{5.6\ cm^2 \times 1}{10,000\ cm^2/m^2} = 0.00056\ m^2$

5.2.3 Area of Templates

Templates may vary from the 12.5 cm2 size standard. Confirm area of template prior to use. All template samples must list the size of the area scraped on the SVF/COC.

$$\frac{12.5\ cm^2 \times "number of templates"}{10,000\ cm^2/m^2} = X\ m^2$$

For example, for one template: $\dfrac{12.5\ cm^2 \times 1}{10,000\ cm^2/m^2} = 0.00125\ m^2$

5.2.4 Reach Wide Chl*a* Composite Calculation

Each type of sample will contribute more or less to the sample as a whole depending upon the number of like-kind samples composited. A weighted average of Chl*a* is determined from the following equation.

∑ [(Rc*Nc)+(Rt*Nt)+(Rh*Nh)]/[(Nc+Nt+Nh)]

Where: R(c,t,h) = Chlorophyll-a lab result from a core, template, and or hoop

N(c,t,h) = Number of each type

For example. Assume: 3 cores were taken (Chloropyll-a = 60.0 mg/m^2

6 templates (Chloropyll-a = 360.0 mg/m^2)

2 hoops (Chloropyll-a = 160.0 mg/m^2)

	# of Type	R
Cores:	3	60.0
Templates:	6	360.0
Hoops:	2	160.0

Therefore: ∑ **[(3X60)+(6X360)+(2X160)]/11**

Reach Weighted Average for Chloropyll-a =   241.8   mg/m^2

6.0 DETERMINATION OF CHL*A* BY HPLC

With DEQ QA officer approval, Standard Methods 10200 H(5) or EPA Method 447.0 may be used if the laboratory confirms the data generated by HPLC compares to the spectrophotometric method. All method specific QA/QC protocols must be followed. For calculation under SM 10200 H(5) use the data from **Table 5-1** of this document.

7.0 REPORTING RESULTS

Results of Chl*a* analyses must be reported in conformance with MT-eWQX specific format. Data providers are required to populate the data in a MT-eWQX EDD. Detailed guidance for populating the MT-eWQX EDD can be found in the MT-eWQX Guidance Manual located on the Web at http://deq.mt.gov/wqinfo/datamgmt/MTEWQX.mcpx.

APPENDIX A STATISTICAL CONFIDENCE IN DEQ'S BENTHIC ALGAE CHL*A* SAMPLING

(Prepared by Michael Suplee, Ph.D., and Rosie Sada)

Definitions:

> **Sample Frame**: Within an assessment reach, a "site", which is a short longitudinal stretch of stream, or a perpendicular transect of a stream/river. For wadeable streams, a site's longitudinal length is defined as 40 times the average wetted width at its midpoint, or 150 m minimum[5]. Sites can have multiple sampling events (i.e., across-time sampling).

> **Population**: The algae growing on or across the surface of the stream bottom within the defined longitudinal area of the site, or across the perpendicular transect.

> **Sampling Unit**: An individual benthic algae sample collected via the template, core, or hoop method, which is subsequently measured for chlorophyll *a* (Chl*a*) and/or ash free dry weight.

Suplee et al. (2009), demonstrate that average benthic algae levels up to about 150 mg Chl*a*/m^2 are considered acceptable by the Montana public majority, whereas average levels of 200 mg Chl*a*/m^2 or more are very undesirable. This algae level threshold is an important component in the assessment of beneficial use attainment, particularly for the western and mountainous regions of the state. As such, it is important to understand the variability and confidence levels associated with DEQ's Standard Operating Procedure (SOP) for sampling benthic algae. This section describes the datasets used, methods, and confidence intervals estimated for DEQ's algal sampling protocols. The purpose of the analysis outlined here was to answer the following question:

1. What is the accuracy and confidence in the average Chla value determined for a sampling event? (The above question is posed given that DEQ's Chla SOP calls for the collection of 11 replicates from a site.)

A.1. DATASETS, AVERAGES, COEFFICIENTS OF VARIATION

All available datasets were collated for the purpose of understanding variability and confidence in DEQ benthic algal sampling methods (**Table A-1**). Data from the Stream Reference Project (STREFPRO) were restricted to sites in western Montana ecoregions, or transitional ecoregions, and did not include algae collected from eastern Montana prairie streams. This is because prairie streams are substantially

---

[5] 40 times the mean wetted width (150 m minimum) was derived from USEPA's method to asses wadeable streams (Lazorchak et al., 1998). Similarly, the USGS NAWQA program uses 20X the wetted width, and Simonson *et al.* (1994) specify 30-35 times the wetted width. These stream lengths have been shown to be sufficiently long to encapsulate key stream characteristics (e.g. depth, substrate type, cover, shading, bank height, etc.) which strongly influence algae and aquatic plant growth. Simonson et al. (1994) find that the aforementioned stream variables could be estimated within ± 5% of the true mean 81-89% of the time using 11 transects spaced along a reach that is 30-35 times the mean wetted width. This longitudinal length was also found to encompass at least three riffle-pool sequences (Leopold et al, 1964).

different ecologically (e.g., they grow considerably more algae naturally, often have fine sediment bottoms, and are commonly intermittent), and can be better assessed for nutrient impairment by examining other factors like dissolved oxygen. The other datasets were from streams or rivers that typically have gravel/cobble substrates and good flow. The University of Montana's Clark Fork River dataset and the PP&L FERC 401 certification dataset were helpful in that many sampling events straddled or exceeded the nuisance threshold value of 150 mg Chl*a*/m^2 identified in Suplee et al. (2009).

The term *sampling event* is used here to describe the sampling of algae (**n** = 10 or 11 replicates) on a given day, as opposed to *site*, since a site may have been sampled repeatedly over time (e.g., monthly during the summer) and we wanted to include these temporal sampling events. Conditions for a sampling event to be used in the analysis were (a) the total number of replicates collected and analyzed for the sampling event was 10 or 11, and (b) the average benthic Chl*a* calculated from a sampling event's replicates was >19 mg Chl*a*/m^2. These conditions were established so that the amount of effort expended for any given sampling event was uniform (i.e., replicate **n** was nearly identical), and so that replicates whose result were non-detects (ND) would not be a substantial part of the dataset. (Sampling events with average algae >19 mg Chl*a*/m^2 usually had none, or at most 1or 2 replicates, as NDs, whereas sampling events whose average algae was <19 mg Chl*a*/m^2 often had several or more NDs.) Also, 19 mg Chl*a*/m^2 is the average algal density we have so far determined for western MT reference streams (*see*: http://deq.mt.gov/wqinfo/nutrientworkgroup/agendasMinutes/2009/july09/NWG_july16-09.pdf), and since we are most interested in variability of algal growth closer to the nuisance threshold, 19 mg Chl*a* /m^2 seemed liked a natural low-end cutoff.

Table A-1. List of Datasets Available for Use in this Analysis.

Dataset	Source	No. Sampling Events in Dataset	Number of Sampling Events Used	Notes
Stream Reference Project (STREFPRO; 2004-2008)	DEQ	72	21	Since these were reference sites, many Chl*a* means <19 mg/m^2
Outstanding Fisheries Project (2006-2007)	DEQ	65	32	
Monitoring – Reassessment (2006)	DEQ	15	0	Data not used. All sampling event averages <19 mg Chl*a*/m^2, or n <<10.
Tabacco & Flint Watershed TMDLs (2007)	DEQ	19	14	
Yellowstone River QUAL2K Model (2007)	DEQ	10	0	Data not used. Sampling event averages <19 mg Chl*a*/m^2, or <<10
Upper Gallatin R. TMDL (2005-2008)	DEQ	?	0	Data not used. 11 transect replicates were composited (no replicates)
PP&L Madison/Missouri Sampling (FERC 401 cert. Compliance)	PP&L	54	47	
Clark Fork River Algae Sampling ('87-'05)	University of MT	328	105	Only used sampling events for which n=10 or 11

The average, standard deviation (**s**), and coefficient of variation (CV; [**s**/average]·100) was calculated from the replicate Chl*a* measurements for each sampling event. The CV is very useful as it allows comparison of any given sampling event's standard deviation to all other sampling event's standard deviations, regardless of whether the sampling event's average algae level was low or high (or in between). After screening the sampling events with the two conditions above (sample size, mean algae >19 mg Chl*a*/m^2), a total of 218 individual sampling events were collated. These 218 sampling events, each comprising 10 or 11 replicate Chl*a* samples each, represent more than 2,200 individual measurements of stream bottom Chl*a*. Template, hoop, and core samples are all represented. They also represent a wide array of stream types and conditions such as rivers with varying amounts of eutrophication (Clark Fork, Madison, and Missouri rivers), low and mid Strahler-order streams with little or no human impacts (STREFPRO), mid-order streams with various degrees of human impacts (TMDL datasets), etc.

A correlation was run between the sampling events' average benthic algal Chl*a* values and their associated CVs (**Fig. A-1**)(Cattaneo and Prairie, 1995). This was carried out to ascertain if there was any clear relationship between the two (e.g., lower algae levels are closely associated with higher replicate CVs, or visa-versa), which could influence subsequent analyses. No clear pattern was noted; the correlation coefficient (r^2) between the two variables was very low, only 0.07. Low average Chl*a* values were associated with about eight unusually high CVs (**Figure A-1**, outside of gray box), but the vast majority of low-Chl*a* averages had CVs typical of the entire range.

A histogram of CVs for the complete dataset was generated (**Figure A-2**). **Figure A-2** shows a clearly defined central tendency for the replicates' CVs (average = 69%). Thus, across all datasets, a typical Chl*a* sampling event comprised of 10 or 11 replicates typically had a replicate CV of 69%.

There were significant differences in CV patterns between the five different datasets (Analysis of Variance; p < 0.05). But each dataset contributed information to the whole that could not be ascertained individually. For example, the Clark Fork River dataset had a lower mean CV than did the other datasets, but provided much information on replicate variability for Chl*a* samples bracketing the nuisance threshold (150 mg Chl*a*/m^2).

The central tendency of the CVs provides a mechanism to estimate an overall confidence and interval width for the SOP algae sampling method (more on this, **Section A.2** below). Consideration must first, however, be given to the fact that the datasets' individual CV patterns differ. Among the 5 datasets, the most variable dataset was the Outstanding Fisheries Project (average replicate CV = 88%), the least variable dataset was the Clark Fork River (CV = 58%), and the dataset falling exactly midway was the Tobacco & Flint TMDL dataset (average replicate CV = 73%). The latter is a DEQ dataset collected following this SOP's methods. Since the central tendency of all datasets is a CV of 69%, and since a DEQ dataset with a mean CV of 73% falls exactly midway among the datasets, 73% is a good estimate of the *typical* replicate variability one would encounter in a *typical* DEQ benthic Chl*a* sampling event. This CVvalue was used to estimate the overall confidence level associated with the method.
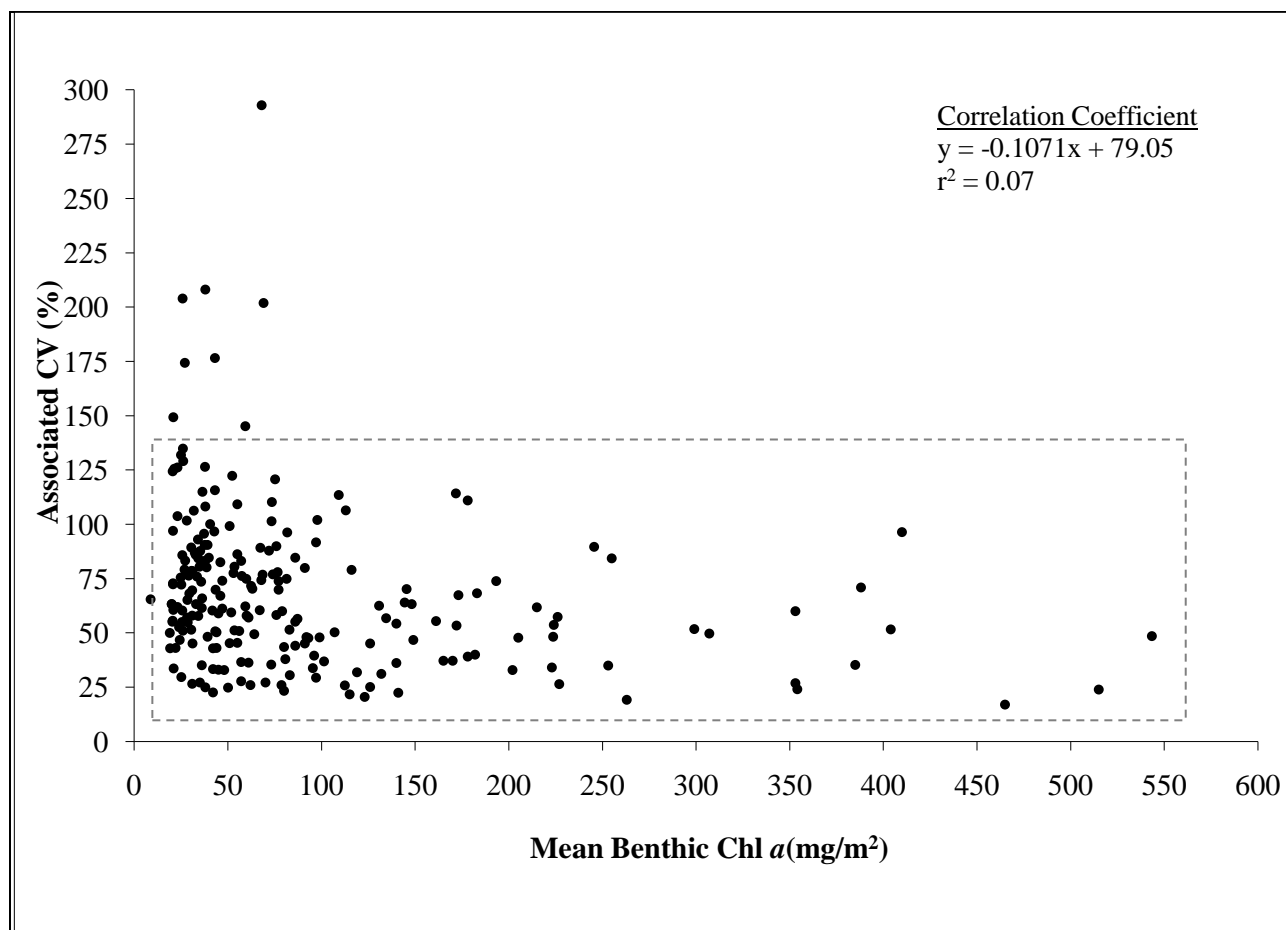
Figure A-1. Scatterplot between mean benthic Chl*a* and Corresponding CVs.
Gray outlined box show that the vast majority of CVs, over a large Chl*a* range, range from about 25 to 120. Regression correlation (R^2=0.07) shows that there is no relationship between the two variables.
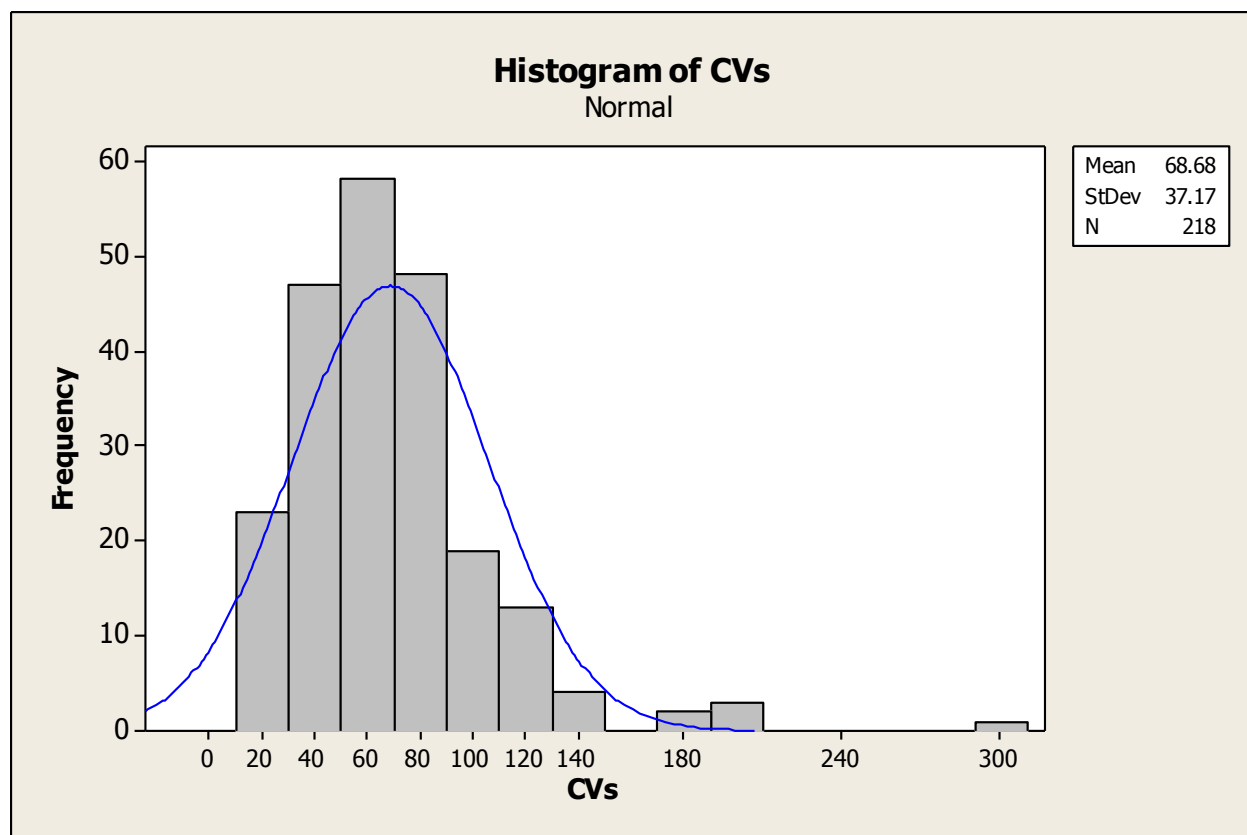
Figure A-2. Histogram of CVs for the 218 benthic Chl*a* sampling events.

A.2. ESTIMATION OF THE CONFIDENCE LEVEL AND INTERVAL WIDTH FOR DEQ'S BENTHIC ALGAE SAMPLING PROTOCOL

The next objective was to estimate the interval width (i.e., the range across which the true population average for a sampling event is likely to fall) and the confidence in that interval. The goal is to be able to make a statement like "with the sampling protocol used, I am usually within ± X of the true benthic algae average Y% of the time".

Two statistical equations (Ott, 1993) and a third statistical approach (bootstrapping; Manly, 2001) were used together to develop these estimates. The two equations used are:

$$\mathbf{n} = \frac{\mathbf{t}^2 \mathbf{s}^2}{\mathbf{d}^2} \qquad\qquad (1)$$

where **n** is the number of sample replicates collected, **t** is the two-tailed critical value of the Student's **t** distribution, **s** is the standard deviation of a sampling event's replicates, and **d** is the pre-selected half width of the desired interval width of the sample mean.

$$\mathbf{x} \pm \mathbf{Z}_{\alpha/2} \, \mathbf{S_x} \qquad\qquad (2)$$

where **x** is the sampling event average Chl*a*, **Z**$_{\alpha/2}$ is the Z-table value having a tail area of $\alpha/2$ to its right, and **S$_x$** is equal to the standard deviation of the sampling event divided by the square root of **n** (**n** =11).

The bootstrapping approach will be discussed momentarily.

Equation 1 is solved iteratively (Thornton et al., 1982), and is commonly used to determine an appropriate sample size for future sampling events, given the variability of a test sample's replicates and the desired interval width. In our case, sample size is about 11, since that is the SOP for Chl*a*. The standard deviation, **s**, needs to be determined for a given sampling event, and **d** needs to be determined based on what degree of precision around the true average is wanted.

Again, our objective was to determine the confidence and accuracy with which our current protocol estimates an average benthic algae level. The most reasonable standard deviation (**s**) to use in equation 1, therefore, is the one associated with the central tendency for all types of streams sampled via these methods (**Figure A-2**). Per our rationale in **Section A.1**, a CV of 73% was used to back-calculate **s**. This is easily done, as **s** = [CV· average]/100. (For example, an 11-replicate sampling event with an average algae density of 114 mg Chl*a*/m^2 will have, at a CV of 73%, an **s** of 83.) The desired interval width also needed to be determined; we believed a precision of ± 20-30% of the mean was reasonable. This is because, at a threshold of 150 mg Chl*a*/m^2 (nuisance threshold), the upper confidence bound for the measured average should fall short of the unacceptable (i.e., 200 mg Chl*a*/m^2) algae level, given a reasonable level of confidence.

Equation 1 was solved iteratively (Thornton et al., 1982) with a fixed **n** = 11, varying confidence levels from the Student's t distribution, and varying values of **d** between 20 and 30% of the average. We found that, with a typical replicate CV of 73%, we are 80% confident that the average benthic Chl*a* measured during a sampling event will be within ± 30% of the true population average. **Stated differently: for a**

typical benthic Chl*a* sampling event that has followed the SOP, DEQ is confident that 80% of the time the <u>measured</u> Chl*a* average will be within ± 30% of the <u>true</u> population average.

It should be noted that if the replicates from a sampling event produce a much lower CV, then greater confidence in the interval width is possible (e.g., 90 or 95% confidence), the interval width can be narrowed, or both. Conversely, sampling events with much higher CVs will estimate the mean Chl*a* with less confidence, or with a wider interval width, or both. The confidence limits estimated in the above paragraph are, therefore, for *typical* Chl*a* sampling events.

Bootstrapping was used to cross-check these results. Bootstrapping is a method that, in the absence of any other knowledge about a population, considers the distribution of values found in a random sample of size **n** from a population to be the best guide to the distribution in the population itself (Manly, 2001). To carry out the bootstrap, test datasets comprised of 11 replicates which produced a particular average algae level (e.g., 150 mg Chl*a*/m^2) were created, each test dataset having a replicate CV of 73%. For any given algae level (say, 150 mg Chl*a*/m^2), the bootstrap program sampled (with replacement) the original 11 replicates, and then generated an average for the re-sampled observations. This was repeated 200 times, i.e., creating 200 generated datasets and 200 corresponding averages. A confidence interval and upper and lower bound for the generated average Chl*a* levels were then calculated. In all cases the results of the bootstrap closely matched the output from equation 2, for both the confidence level and the interval width.

Finally, we examined the consistency of averages calculated from field duplication efforts. These involved the collection of 11 replicates along a site, labeled "Duplicate 1", followed immediately by the collection of another 11 replicates, labeled "Duplicate 2". The Duplicate 2 samples were collected either 1 m up- or downstream of the Duplicate 1 collection points, or were collected via one of the site's alternative longitudinal sampling patterns (see **Section 1.2.5** and **Figure 1-0**, this SOP). There is no way to know which (if either) of the two duplicate values is the "true" Chl*a* average, therefore the average of all the replicates (n=22) at a site was assumed to be the true site average. The % difference between the collective average and each duplicate's average was then evaluated (i.e., did duplicate 1 and duplicate 2 each fall within 30% of the all-replicates average). Results are shown in **Table A-2**.

Table A-2. Field Duplicated Sampling Events for Benthic Algal Chl*a*.
Values are the average benthic Chl*a* calculated seperately for two sampling events (N = 11 reps each), carried out at the same stream site, same day.

Site	Benthic Algal Chl*a* (mg/m^2)		CVs (%)		Duplicates' average Chl*a* within expected range*?
	1st Duplicate	2nd Duplicate	1st Duplicate	2nd Duplicate	
Rock Creek (Y02ROCKC01)	21	21	126	149	Yes
Beaver Creek (M09BEVRC05)	51	67	99	61	Yes
Little Thompson River (C13LTTPR40)	32	23	106	104	Yes
Swamp Creek (C13SWPCR20)	12	11	62	61	Yes
E. Gallatin River (site EG10)	136	130	n/a	n/a	Yes
E. Gallatin River (site EG13)	54	161	n/a	n/a	No
Armells Creek (M31ARMLC07)	58	65	70	89	Yes
Middle Fork Judith River (M22JUDSF01)	31	36	80	96	Yes
Moose Creek (M03MOOSEC04)	26	27	112	132	Yes
Eagle Creek (M10EAGLC01)	79	96	83	79	Yes
Box Elder Creek (Y26BOXEC08)	22	31	74	78	Yes

*Compared to ± 30% of the whole dataset average, i.e., all replicates collected for the 1st duplicate and the 2nd duplicate. Note: In some sites the templates, hoops or cores were composited and so a CV could not be calculated. Shown as n/a in the 'CVs (%)' columns.

Overall, the duplicates' averages all fell within ± 30% of the corresponding overall average, with one exception. Nine of eleven successful duplications is 91% success; this is in fact superior to the calculated statistical confidence level (80%) of the method. These data therefore show that actual field duplication will produce results equal to or better than our statistically determined expectations derived from replicate variability.

APPENDIX B

Aquatic Plant Tracking Form

Waterbody: _____

Site Visit Code: _____

Date: _____

Visit No.: _____

Transect	Position on stream (R, L, C)	Sampling Method*	*For templates*: Number of templates collected	*For templates*: Number of GF/F filters on which the sample has been collected	Dominant Algae†	For Hoop Samples (check one) 100% algae ←——→ 100% macrophytes					Moss (% bed cover in a hoop)‡
						100/0	75/25	50/50	25/75	0/100	
A											
B											
C											
D											
E											
F											
G											
H											
I											
J											
K											

* H-hoop; C-core; T-template, N-None collected.

† F-filamentous; M-algal film on a mud surface; R-non-filamentous algae on rock surfaces

‡ If moss is the dominant plant at a locale, visually estimate proportion of streambed cover by moss within area of 1 hoop.

REFERENCES

APHA. 1998. Standard Methods for the Examination of Water and Wastewater, 20th ed., American Public Health Association.

Arar E. 1997. EPA Method 446.0 - *In Vitro* Determination of Chlorophylls a, B, C_1 + C_2 and Pheopigments in Marine And Freshwater Algae by Visible Spectrophotometry, Revision 1.2. Washington, D.C.: U.S. Environmental Protection Agency.

Cattaneo, A. and Y. Prairie. 1995. Temporal Variability in the Chemical Characteristics Along the Rivi`Ere De L'Achigan: How Many Samples Are Necessary to Describe Stream Chemistry? *Canadian Journal of Fish and Aquatic Sciences.* 52828-835

Lazorchak, James M., Donald J. Klemm, and David V. Peck. 1998. Environmental Monitoring and Assessment Program – Surface Waters: Field Operations and Methods For Measuring the Ecological Condition Of Wadeable Streams. Washington D.C.: U.S. Environmental Protection Agency.

Leopold, L. B, M. G. Wolman, and J. P. Miller. 1964. Fluvial Processes in Geomorphology, St. Louis, MO: Freeman Press.

Manly, B. F. J. 2001. Randomization, Bootstrap and Monte Carlo Methods in Biology, 2nd ed., Boca Raton, FL: Chapman & Hall.

Ott, R. Lyman. 1993. An Introduction to Statistical Methods and Data Analysis - Fourth Edition, Belmont, CA: Duxbury Press.

Sartory D.P. and Grobbelaar J.U. 1984. Extraction of Chlorophyll a From Freshwater Phytoplankton for Spectrophotometric Analysis. The Hague: Dr. W. Junk Publishers. Hydrobiologica 114.

Simonson, T. D., J. Lyons, and P. D. Kanehl. 1994. Quantifying Fish Habitat in Streams: Transect Spacing, Sample Size, and a Proposed Framework. *North American Journal of Fisheries Management.* 14606-615

Suplee, M., V. Watson, M. Teply, and H. McKee. 2009. How Green Is Too Green? Public Opinion of What Constitutes Undesirable Algae Levels in Streams. *Journal of the American Water Resources Association.* 45123-140

Suplee, M., V. Watson, A. Varghese, and J. CLeland. 2008. Scientific and Technical Basis of the Numeric Nutrient Criteria for Montana's Wadeable Streams and Rivers. Helena, MT: MT DEQ Water Quality Planning Bureau.

Suplee, Michael W., Vicki Watson, and et al. 2006. An Experiment to Determine the Efficiency of Measuring Chlorophyll a in Stream Sediment Samples Using Spectrophotometric Methods. Montana Department of Environmental Quality.

Thornton, K. W., R. H. Kennedy, A. D. Magoun, and G. E. Saul. 1982. Reservoir Water Quality Sampling Design. *Water Resources Bulletin.* 18471-480